

Symbiodinium Distribution Patterns in Millepores in the Caribbean: South Water Cay, Belize and San Salvador, The Bahamas

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Abstract Symbiosis with *Symbiodinium* (Dinoflagellata, Chromalveolata) plays a major role in the survival of numerous coral species. Different *Symbiodinium* clades provide their hosts with different physiological advantages. However, much of the research concerning coral-*Symbiodinium* associations has focused on scleractinian corals, while reef-building Millepores (fire coral) have mostly been ignored. We examined the Millepore-*Symbiodinium* relationship at two thermally different sites in the Caribbean: San Salvador, The Bahamas and South Water Cay, Belize. Our results indicate that there is a difference in symbiont dominance between the two sites. Millepores residing in the cooler sea surface temperatures (SST) of The Bahamas are *Symbiodinium* clade B dominant (100%) and are not showing signs of bleaching. However, Belize Millepores experiencing warmer SST are mostly *Symbiodinium* clade A dominant (72%) and are not showing signs of bleaching. Some clade B dominant Belize Millepores (28%) are showing signs of bleaching. These findings are corroborated by a reversal in symbiont dominance of Belize Millepores residing at deeper depths; most are *Symbiodinium* clade B dominant (80%) and none of these colonies showed signs of bleaching. Our results appear to correlate with the temperature difference between the two sites, and suggest Belize Millepores may be experiencing thermal stress events more frequently. The relationship between specific *Symbiodinium* clades and their coral hosts may provide the coral with a mechanism to cope with increased thermal stress due to global warming.

Keywords *Symbiodinium*, Millepore, Clade dominance

1. Introduction

Coral reefs host more than 25% of all known marine species [1], but recent estimates have projected a 70-90% decline in coral reefs by 2050 due to climate change [2]. The loss of coral cover not only decreases biodiversity, but also has negative impacts on local and global economies. Reefs generate revenue by supporting coastal development through the fishing and tourism industries, which in turn creates jobs. Coral reefs have an estimated net worldwide economic value of over \$30 billion USD per year [1]. Small tropical nations especially rely on the tourism trade. In Bermuda, for example, 12% of the yearly gross domestic product (GDP) depends on local reefs [3]. In Belize, coral reefs contribute up to 15% of the GDP [4]. Coral reefs also comprise a major food source for over 500 million people around the world [1]. Therefore, gaining a better

understanding of the impact of climate change induced stressors on coral reefs is of the utmost importance.

Global warming has been linked to coral bleaching, and is causing devastating effects on coral reefs and the biodiversity they support [5]. Coral bleaching usually occurs when sea surface temperatures (SST) reach 30°C or above for 3 to 4 weeks, which in most cases only requires a 1 to 2°C water temperature increase above ambient temperatures [6]. Since the 1980s, there have been three pan-tropical bleaching events (1998, 2010, and 2015-2016) along with numerous regional mass bleaching events, resulting in coral death and decreased coral cover [7]. Hughes et al. [7] noted that with each successive bleaching event, fewer reefs escaped unscathed. During the 1998 bleaching event, 45% of the surveyed reefs survived with no bleaching, but during the 2016 event, only 9% of the surveyed reefs escaped with no bleaching [7]. Bleaching is caused by the loss of symbiotic algae (*Symbiodinium*) that reside in the corals' tissue [8]. *Symbiodinium* can provide up to 95% of the corals' nutritional needs, so losing

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symbionts during bleaching weakens the corals and makes them more susceptible to disease and future climatic events [9].

Symbiodinium, commonly known as zooxanthellae, are a group of photosynthetic dinoflagellates that form symbiotic relationships predominantly with reef-building corals, as well as with a wide range of marine invertebrates including Porifera, other Cnidarians, Platyhelminthes, and Mollusca. The vibrancy of coral colors comes from the photosynthetic pigments of the algae they house [8]. There is a high degree of diversity among *Symbiodinium*, and they are genetically classified into nine clades, A through I, with numerous subclades [10]. *Symbiodinium* tend to have a high specificity for their coral hosts, and hosts exhibit a genetic preference for certain clades of *Symbiodinium*. For example, Scleractinia (stony) corals form symbiotic interactions with *Symbiodinium* clades A, B, C, D, and F, while Hydrozoan (fire) corals predominantly associate with *Symbiodinium* clades A, B, and C [10]. Although *Symbiodinium*-coral interactions are specific, corals that form associations with *Symbiodinium* can contain either one clade, or one dominant with multiple background clades [11]. Baker and Romanski [12] found that 72% of the coral species they examined had multiple clades and subclades, indicating that *Symbiodinium*-host associations are flexible.

Corals that host *Symbiodinium* grow on shallow reefs where there is more light, and temperatures are already close to the tipping point for bleaching. *Symbiodinium* are located within organelles called symbiosomes, and photosynthetic products and other nutrients are passed between the host and the symbiont through the symbiosome membrane [13]. The coral, in return, provide the *Symbiodinium* with protection and exposure to light [14]. However, numerous environmental factors can stress the relationship resulting in a breakdown of symbiosis, loss of the symbionts, and, ultimately, coral bleaching [11]. The physiological mechanisms that lead to the selective loss of symbionts and bleaching may be due to the symbiont's release of reactive oxygen species (ROS) during photosynthesis [15].

Under non-stressful environmental conditions, ROS produced during photosynthesis are mitigated by antioxidants [16]. However, increased irradiance levels lead to increased ROS which overwhelms the symbionts' photosystems' ability to scavenge ROS [17]. Wietheger *et al.* [15] measured the production of ROS in cultured *Symbiodinium* clades A1, B2, and F1 under thermal and oxidative stress conditions using fluorescent probes, and observed significant increases in ROS in all clades. Oxidation caused by ROS has been postulated to destroy the symbiosome membrane, damage host cell proteins and membranes, and cause the loss of symbionts, resulting in bleaching [18]. However, the hypothesis that ROS are responsible for bleaching is controversial. Nielsen *et al.* [19] found it unlikely that oxidative stress to host cells was the cause of coral bleaching in *Pocillopora damicornis* (cauliflower coral).

The stress that results in bleaching may provide the opportunity for coral to adapt to changing environmental conditions by allowing coral to alternate between specific symbionts that confer physiological advantages to their hosts. The Adaptive Bleaching Hypothesis (ABH) proposes that stress events provide an opportunity for the host to be repopulated with different symbionts that are better adapted to help the coral survive environmental change [16]. This exchange occurs in one of two ways: either switching (recruitment of new symbionts from the environment) or shuffling (changes in the dominant symbiont present in a coral colony) [20,21,22]. Switching or shuffling symbiont clades allow corals to "select" the symbiotic relationship best suited to cope with environmental stressors, and thus allow corals to better adapt to their current environment.

Since it was first proposed, much evidence has supported the ABH. Baker *et al.* [23] surveyed reefs in Kenya, Mauritius, the Red Sea, Persian Gulf, and Panama during the pre-stress (1995), stress (1997), and post-stress (2001) periods of the 1997-1998 El Niño event. They found that corals that originally contained C as their dominant clade and shuffled clade D from a background clade to a dominant clade were less affected by the increased temperatures than those coral that remained C dominant. They concluded that clade D is more thermally tolerant and the symbiosis could survive periods of thermal stress [23]. Berkelmans and van Oppen [20] found that *Acropora millepora*, a common hard coral in the Indo-Pacific, could increase their thermal tolerance by 1-1.5°C by changing the dominant *Symbiodinium* clade from C to D.

There is also evidence that *Symbiodinium*-host interactions may not be as flexible as once believed, but instead are governed genetically, not environmentally. Poland and Coffroth [24] took the aposymbiotic offspring of the Caribbean octocoral (*Briareum asbestinum*) from a parental colony that associated with *Symbiodinium* clade B184 and transplanted the juveniles to a similar habitat where the existing *B. asbestinum* colonies associated with *Symbiodinium* clade B178. Transplanted juvenile coral polyps formed symbiotic relationships with many different clades, but after about 4 years, the offspring had associated almost entirely with the dominant clade of the parent colonies (B184), despite the transplant location and the resident *B. asbestinum* colonies containing a different dominant clade [24]. They concluded that the genetics of both the symbiont and the host coral determine the symbiont diversity, not the environmental conditions. Palumbi *et al.* [25] showed that transplanted *Acropora hyacinthus* (bush coral) could adapt to thermal stress with no change to their symbiont population. Therefore, the ABH alone may not explain coral's adaptive mechanism to cope with thermal stress in all situations.

The Caribbean offers an ideal environment to study thermal stress in host-symbiont associations because from 1955 to 2016, the Caribbean has experienced an average temperature increase of 0.24°C per decade [26]. These warming trends are sufficient to push coral outside their

range of thermal tolerance. There has been extensive research on *Symbiodinium*-scleractinian (stony coral) interactions in these conditions, but the ubiquitous reef-building Millepores (fire coral) have largely been ignored [27]. *Millepora*, a genus of hydrozoan coral, are common encrusting corals found throughout the Caribbean [28]. Traditionally, Caribbean Millepores are classified into two species, *Millepora alcicornis* and *Millepora complanata*, based on morphology [28]. *M. alcicornis* are thinly encrusting sheets and branches found mostly at deeper depths with less wave action and *M. complanata* are thick rigid blades found mostly at shallower depths [28]. However, the Millepores exhibit phenotypic plasticity resulting in the presence of many intermediate growth forms whose taxonomic status is uncertain [27,28,29,30,31].

We have focused our attention on *Symbiodinium*-Millepore symbiosis at two study sites where yearly SST is different: San Salvador, The Bahamas (22-28°C) [32] and South Water Cay, Belize (26-31°C) [33]. Since coral bleaching generally occurs slightly above 30°C, the corals at the Belize site may be under greater thermal stress due to heightened irradiance levels, whereas the corals residing in Bahamian reefs may be less likely to cross the bleaching threshold, resulting in a possible difference in symbiont population.

In a preliminary study with a small sample size, Samayoa et al. [34] found that clade B might offer Millepores a physiological advantage in the cooler SST of the Bahamas, while clade A might be more beneficial to Millepores in Belize with warmer SST. Here we report that *Symbiodinium* clade dominance in Millepores differs significantly between cooler SST in reefs around San Salvador, The Bahamas compared to warmer SST in reefs close to South Water Cay, Belize. Belize Millepores residing in shallow patch reefs are mostly clade A dominant. However, colonies collected from deeper reefs in Belize are mostly clade B dominant, indicating a change in dominance with regard to depth.

Determining whether certain clades of symbionts offer Millepores a thermal physiological advantage will provide insight into future conservation efforts for corals in the Caribbean and worldwide.

2. Methods and Materials

2.1. Collection and Preparation of Millepore Samples

Millepores used in this study were collected via snorkeling or SCUBA from 2009-2018 on reefs on the west side of San Salvador, The Bahamas (24.1°N, 74.4°E) and from 2013-2019 along the barrier of the South Water Cay Marine Reserve (SWCMR) near South Water Cay, Belize (16.8°N, 88.1°E) (Figure 1). San Salvador is located on the eastern edge of The Bahamas Island Archipelago. Collection sites were shallow (1-5 m) patch reefs located at Lindsay Reef, Rocky Point Reef, Grotto, and French Bay. South Water Cay,

which is part of Belize's largest marine reserve, covers an area of 117,878 acres and is located approximately 15 miles southeast of Dangriga. Millepore samples were collected from shallow (1-5 m) sites around SWCMR located at Whales Shoal, Aquarium, Angel's Reef, Coral Garden, Tobacco Cay, IZE Reef, and Curlew Reef and deeper (5-24 m) sites at Carrie Bow Fore Reef, Coral Garden, and Aquarium Wall.

Various morphotypes of Millepores (*M. alcicornis*-branched, *M. complanata*-bladed, and intermediate morphologies) were randomly collected at shallow locations by removing a small piece (approximately 4 sq. cm) of coral. Only *M. alcicornis* samples resided at deeper locations. Samples were transported from collection sites to the lab and held for no more than two days prior to DNA isolation.

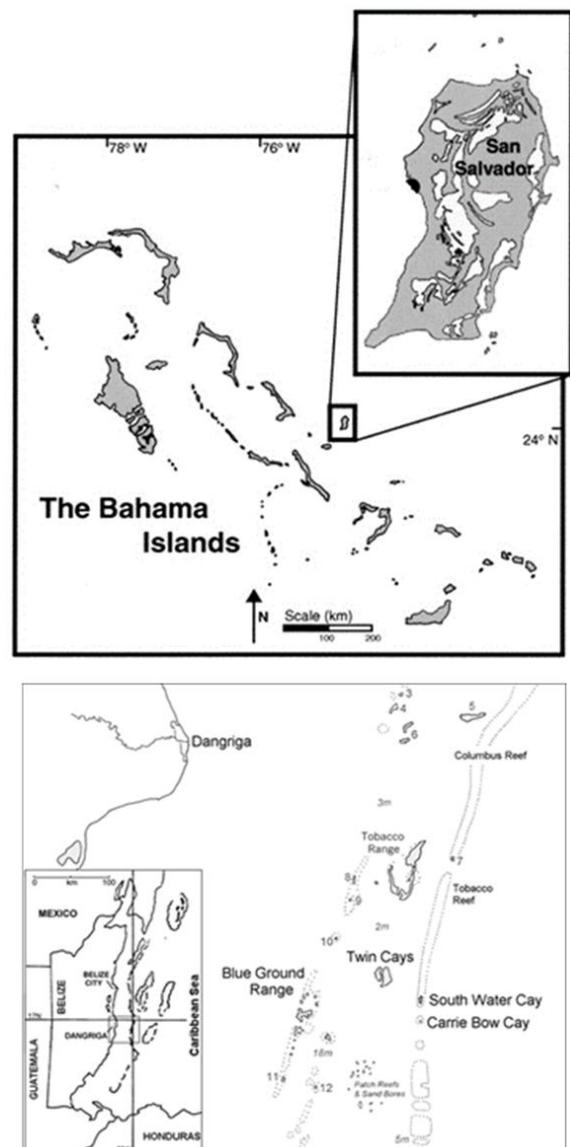


Figure 1. Map locations of collection sites: (Top) San Salvador, The Bahamas (<http://palaos.sepmonline.org/content/16/4/372/F1.large.jpg>) and (Bottom) South Water Cay, Belize (<http://cbc.riocean.com/location.htm>)

2.2. *Symbiodinium* DNA Isolation

Symbiodinium DNA was isolated using the procedure of Samayoa *et al.* (34). Millepore samples were repeatedly blasted with L buffer (100 mM EDTA, 10 mM Tris, pH 7.6) from a 50cc syringe. Removed coral tissue was centrifuged at 3,500 rpm for 10 minutes and the resulting pellet was washed in 10 mL of L buffer and recentrifuged. The tissue pellet was resuspended in 900 μ L of L buffer and macerated manually with a tissue homogenizer. The homogenate was centrifuged at 13,000 rpm for 10 minutes and the pellet resuspended in L buffer. *Symbiodinium* were lysed with SDS (1% w/v) and incubated at 65°C for 60 minutes. Pro K (0.5 mg/mL) was added and the lysate was incubated at 37°C for a minimum of 6 hours. NaCl (0.8 M) and CTAB (1% w/v) were added and the samples incubated at 65°C for 30 minutes. Nucleic acids were ethanol (70%, v/v) precipitated (twice) in sodium acetate (0.3M, pH 5.2) and immediately centrifuged. The pellet was resuspended in dH₂O.

2.3. Quantitative PCR

Four clade-specific *Symbiodinium* primer pairs [35,36] targeting specific regions of rDNA were used (Table 1). The 20 μ L qPCR reaction contained 10 μ L of Power SYBR Green Mastermix (Applied Biosystems), 75 nM of clade-specific primer pairs, and 200 ng of *Symbiodinium* DNA. Amplifications were performed and analyzed on an Applied Biosystems StepOnePlus Real-Time PCR system. The qPCR profile consisted of an initial denaturation of 95°C for 10 minutes followed by 40 cycles of 95°C for 30 seconds, 58.5°C for 30 seconds and 72°C for 30 seconds. In order to determine if amplification signals represented the amplicon or primer dimers, at the end of each run a melt curve was generated starting at 55°C and increasing the temperature by

0.5°C every 5 seconds until a temperature of 95°C was reached.

The cycle-threshold (C_T) was fixed at a value of 0.1 to allow for the comparison of C_T values between runs. Each run contained duplicate reactions for each clade as well as negative (no-template) controls. C_T values for each duplicate reaction were averaged.

Table 1. Clade-specific qPCR Primer Pairs Used for Amplifying Portions of Multi-copy rDNA in *Symbiodinium* Clades A, B, C and D [35,36]

| | Primer Pair | Target |
|---|--|--------------------|
| A | 5'CCTCTTGGACCTTCCACAAC3' 5'GCATGCAGCAACTGCTC3' | ITS1-5.8S -ITS2 |
| B | 5'GTCTTTGTGAGCCTTGAGC3' 5'GCACACTAACAAAGTGTACCATG3' | LSU-28S |
| C | 5'AAGGAGAAGTCGTAACAAGGTTTCC3' 5'AAGCATCCCTCACAGCCAAA3' | ITS1 |
| D | 5'AAGGAGAAGTCGTAACAAGGTTTCC3' 5'CACCGTAGTGGTTCACGTGTAATAG3' | ITS1 |

3. Results

The same sample set from numerous shallow (1-5 meters) reefs was used to examine the effects of geographic location, morphology, year and temperature, and individual reef location. The individual reefs were in close proximity to each other, and no significant trends were observed between reefs (data not shown).

3.1. Geographic Location

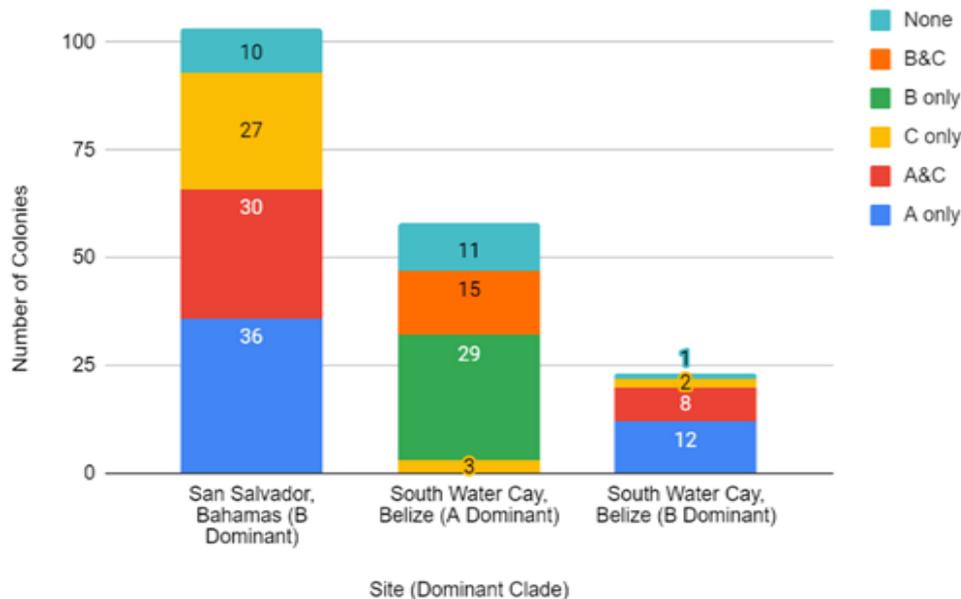


Figure 2. Comparison of the dominant and background *Symbiodinium* clades present in Millepores collected from reefs around San Salvador, The Bahamas (N=103) and on reefs around South Water Cay, Belize (N=81). The x-axis shows the dominant clade residing in Millepore colonies analyzed and the numbers on the graph represent the background clades in each dominant clade group (y-axis)

A total of 184 Millepore colonies were examined for *Symbiodinium* clades A, B, C, and D from numerous reefs surrounding two study sites, South Water Cay, Belize (N=81) (2013-2017) and San Salvador, The Bahamas (N=103) (2009-2018) (Figure 2). In the Bahamas, *Symbiodinium* clade B was the dominant clade in all Millepores examined (100%). In Belize, 72% (N=58) of the Millepores were clade A dominant and 28% (N=23) were clade B dominant. Belize *Symbiodinium* B dominant Millepores seemed to have a greater tendency to bleach (N=5) than *Symbiodinium* A dominant Millepores in the same areas (Figure 3). No Millepore bleaching was observed in The Bahamas.

The majority of Millepores examined in The Bahamas and Belize contained more than one *Symbiodinium* clade. In The Bahamas, background clades were found in 90% (N=93) of the colonies. Of the colonies that contained background clades, 64% (N=66) contained clade A either alone or in

combination with clade C (Figure 1); low abundance levels of clade C were found in 55% (N=57) of the colonies. In Belize, background clades were found in 85% (N=69) of the colonies. Of the *Symbiodinium* A dominant Millepores with background clades (N=48), 92% (N=44) contained clade B either alone or in combination with clade C (Figure 2); low abundance levels of clade C were found in 38% (N=18) of the colonies. Of the clade B dominant Millepores with background clades (N=22), 91% (N=20) contained clade A either alone or in combination with clade C (Figure 2); low abundance levels of clade C were found in 45% (N=10) of the colonies. *Symbiodinium* clade D was not found in any of the samples, and has not been found in Hydrozoan Millepores previously [10,37]. The presence of background clades in the examined colonies indicates a potential flexibility in the relationship between the *Symbiodinium* and *Millepora*.



Figure 3. Millepores residing in Belize. (A) Bleached *Millepora* containing clade B *Symbiodinium* as the dominant symbiont and (B) healthy *Millepora* containing clade A *Symbiodinium* as the dominant symbiont

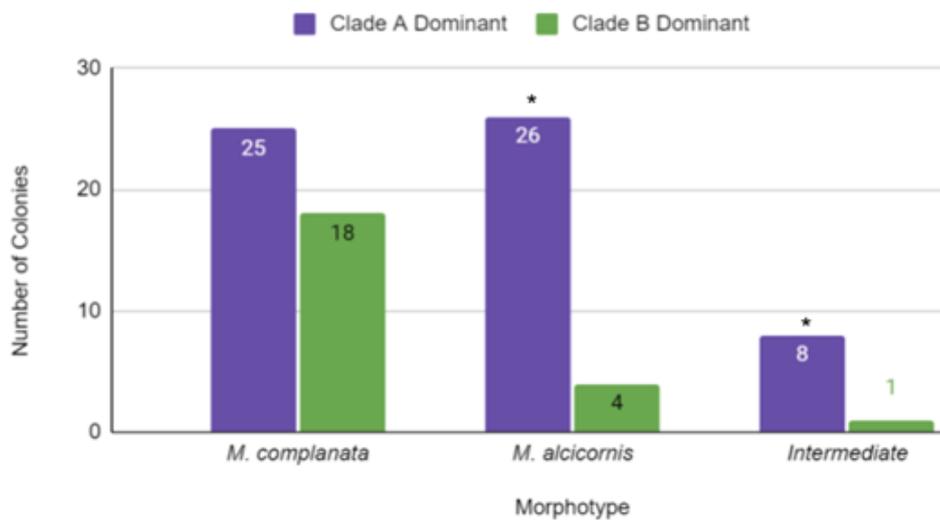


Figure 4. Comparison of the dominant *Symbiodinium* clades present in different morphotypes (growth forms) of Millepores from Belize. The graph shows the number of colonies tested for each morphotype that was either *Symbiodinium* clade A or B dominant (N=82). The asterisk indicates a statistically significant difference between the two *Symbiodinium* clades using a chi-square test for goodness of fit and assuming the two values should be equal

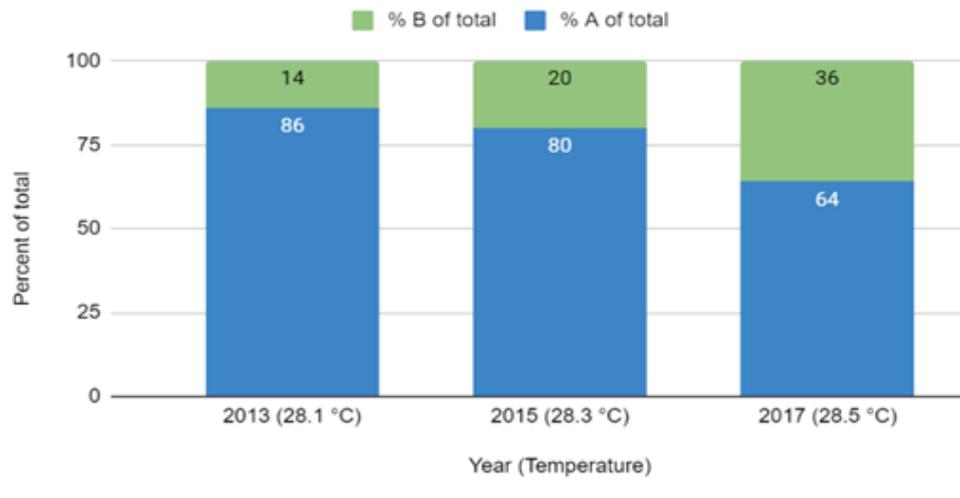


Figure 5. Comparison of *Symbiodinium* clades A and B dominance in Belize collected during three different years (2013, 2015 and 2017). The graph shows the percentage of Millepore colonies that were either *Symbiodinium* clade A or B dominant during a particular year, as well as the average temperature for that year

3.2. Morphology

The dominant *Symbiodinium* clade present in Millepore morphotypes in Belize was also examined (Figure 4). *M. complanata* (bladed morphotype) typically resides in the surf zone while *M. alcicornis* (branched morphotype) and the intermediate morphotype reside 1-5 m from the surface [29]. *M. complanata* (N=43) were split 58% (N=25) clade A dominant and 42% (N=18) clade B dominant. *M. alcicornis* (N=30) were 87% (N=26) clade A dominant and 13% (N=4) clade B dominant. The intermediate morphotypes (N=9) were 89% (N=8) clade A dominant and 11% (N=1) *Symbiodinium* B dominant.

The background clades were also examined, but no significant trends were observed with regard to morphotype (data not shown).

3.3. Year vs Temperature

The yearly dominance of *Symbiodinium* clades in Belize Millepores was compared to year collected and temperature (Figure 5). The average SST for 2013 in South Water Cay, Belize was 28.2°C with a total of 35 days above 30°C [38]. A total of 25 Millepore colonies were analyzed, 76% (N=19) were *Symbiodinium* clade A dominant and 42% of the clade A dominant corals had a clade B background. Of the 24% clade B dominant corals, all had clade A background. Hughes *et al.* [7] reported that between 2015–2016 record temperatures triggered a global pan-tropical episode of coral bleaching. The average SST in 2015 was 28.3°C with a total of 71 days above 30° [38]. A total of 11 Millepore colonies were analyzed, 73% (N=8) were *Symbiodinium* clade A dominant and 87% of the clade A dominant corals had a clade B background. Of the 13% clade B dominant corals, all had clade A background. In 2017, *Symbiodinium* clade A dominance in Belize Millepores had decrease to 56% (N=33) of the 59 colonies analyzed and 91% of the clade A dominant corals had a clade B background. Of the 44% clade B

dominant coral, 84% had clade A background. However, the average SST in 2017 was 28.5°C with 118 days above 30°C and 76 of those days were consecutive [38].

3.4. Deep Samples

An additional survey was conducted on Belize *M. alcicornis* colonies collected from reefs 5-24 meters deep. A total of 35 colonies were examined; 80% (N=28) were *Symbiodinium* clade B dominant and 20% (N=7) were *Symbiodinium* clade A dominant.

The background clades for the deep samples were also examined. Deep *Symbiodinium* clade B dominant colonies contained background clades in 89% (N=25) of the colonies; of those, 96% (N=24) contained clade A as a background clade either alone or in combination with clade C. All the deep *Symbiodinium* clade A dominant colonies contained clade B as a background clade either alone or in combination with clade C.

4. Discussion

Changes in the abundance of symbiont populations may provide a mechanism through which coral hosts can better cope with changing environmental conditions. Different symbionts confer different physiological advantages to the host [39,40]. Berkelmans and van Oppen [20] showed that *Acropora millepora* (branching stony coral) could increase their thermal tolerance by 1-1.5°C by changing their symbiont population from *Symbiodinium* clade C dominant to *Symbiodinium* clade D dominant. Baker *et al.* [23] found that *Pocillopora* colonies had altered their symbiont associations after the 1997-1998 El Niño Southern Oscillation (ENSO) event to a more thermally tolerant pairing with *Symbiodinium* clade D, and concluded that it was possible for coral to revert back to their original pairings given enough time without additional thermal stress events.

These studies provide support for the Adaptive Bleaching Hypothesis (ABH), which proposes that thermal stress events allow hosts to shuffle or switch their symbiont populations to adapt to environmental change [16,41].

The key result of this study is the difference in *Symbiodinium* clade dominance between San Salvador, The Bahamas, with a yearly SST range of 22-28°C [32] and South Water Cay, Belize, with a yearly sea surface temperature (SST) range of 26-31°C [33]. In 2017, South Water Cay had a total of 118 days of SSTs above 30°C, of which 76 were consecutive [42]. The Millepores in The Bahamas are all *Symbiodinium* clade B dominant, and the Millepores in Belize are 72% *Symbiodinium* clade A dominant and 28% *Symbiodinium* clade B dominant. Of the Millepore samples examined, the only colonies that showed signs of bleaching were Belize *Symbiodinium* clade B dominant (N=5) (Figure 3). No Belize clade A dominant colonies or Bahamian clade B dominant colonies examined showed signs of bleaching. The difference in dominant symbiont population between the two sites may be due to the temperature differences between the two geographic locations and suggests that clade A may provide Millepores in Belize with a mechanism to cope with the higher temperatures and irradiance they experience.

Other researchers have observed similar patterns of changes in symbiont populations in the Caribbean. Venn et al. [43] found that, in *Condylactis gigantea* (ball anemone), *Symbiodinium* clade A predominates at inshore and near shore sites that are more thermally variable (15-32°C) and that clade B is the dominant clade at offshore sites with more uniform temperatures (18-28°C). They also observed that anemones that form dominant clade symbioses with clade B are more likely to bleach at SST above 32°C, but anemones that associate with clade A as their dominant clade did not show signs of bleaching. Reynolds et al. [44] used serial irradiation pulse (SIP) fluorescence analysis to determine that *Symbiodinium* clade A exhibits enhanced capabilities for alternative photosynthetic electron-transport pathways and the ability to undergo pronounced light-induced dissociation of antenna complexes from photosystem II reaction centers. They concluded that these facets of *Symbiodinium* clade A allow them to promote the survival of most cnidarian hosts under thermal stress conditions and provide the host with resistance to bleaching. Kemp et al. [45] examined *Orbicella faveolata* (mountainous star coral) before, during, and after a bleaching event in Puerto Morelos, Mexico and found that colonies that contain *Symbiodinium* clade A3 as their dominant clade prior to a thermal stress event were more resistant to bleaching than colonies that contain clade B17 or C7 as their dominant clades. They concluded that this difference is due to clade A3 maintaining a higher quantum yield of photosystem II than clade B17 or C7.

The observations of Venn et al. [43], Reynolds et al. [44], and Kemp et al. [45] are in line with our study which shows an increase in *Symbiodinium* clade A dominance at more thermally stressed sites, as well as the possibility of an increased likelihood of bleaching in Belize clade B dominant

Millepores. The higher thermal tolerance of *Symbiodinium* clade A appears to allow Millepores to cope more effectively with higher temperatures and irradiance in Belize.

Phenotypic plasticity of the Millepores makes morphological characteristics problematic as a tool in taxonomic identification of Millepores [46]. Tepper et al. [29] showed that *M. alcicornis* and *M. complanata* are phenotypically plastic and may, in fact, form a single species-complex that cannot be distinguished by morphological characteristics. Although *Symbiodinium* clade A dominance appears to be significantly more prevalent in *M. alcicornis* (87%) and intermediate growth forms (89%) than clade B dominance (Figure 4), no definitive interpretations of this difference can be made because of the role that the environment plays in phenotypic plasticity of *Millepora* [27,28,29,30,31].

A single Millepore colony is able to host multiple clades of *Symbiodinium* at the same time, which provide Millepores with the capability to alter their symbiont populations to cope with changing environmental conditions. In The Bahamas, all the colonies examined were *Symbiodinium* clade B dominant and 64% of all the colonies examined contained *Symbiodinium* clade A as a background clade. In Belize, *Symbiodinium* clade B was found in 75% of all clade A dominant Millepores, and *Symbiodinium* clade A was found in 87% of all the clade B dominant Millepore colonies. LaJeunesse [37] and Baker [11] noted that although host-symbiont associations are specific, there is flexibility within those specific associations. Fay and Weber [48] found that mixed infection (symbiosis with multiple clades of *Symbiodinium*) occurs in over 30 genera of scleractinian corals. *Symbiodinium* clade A's prevalence as a background clade in *Symbiodinium* clade B dominant Millepores, both in Belize and The Bahamas, suggests that corals have the ability to shuffle the more thermally tolerant clade A into a dominant position, should the warming trend (0.24°C per decade) in the Caribbean continue and environmental conditions necessitate it [26]. The prevalence of *Symbiodinium* clade B as a background clade in clade A dominant Belize Millepores suggests that it is possible that the current clade A dominant Millepores were clade B dominant and have shuffled clade A into a dominant position to cope with the higher temperatures they experienced. Additionally, these clade A dominant colonies have the capability to shuffle back to clade B dominance if the temperature stress is alleviated.

Although clade C, as a background clade, is almost always the least abundant of symbiont clades present in Millepores, it is possible its presence is an advantage for its host. Baker et al. [49] showed that clade C has a higher rate of nitrogen acquisition and an enhanced ability to confer photosynthetic products to the host.

We also examined how symbiont populations fluctuated over time and with different SSTs in Belize. A pan-tropical bleaching event occurred during our study from 2015 to 2016 [7]. Leading up to the bleaching event, in 2013 and 2015, the Millepores were 76% and 73% clade A dominant

respectively (Figure 5). However, although the SST rose slightly in 2017, this year could have still been a period of recovery for the Millepores which led to the decrease in clade A dominance (56%). High percentages of coral colonies with clade B backgrounds, may account for this shuffle in dominance. Thornhill *et al.* [47] observed a similar symbiont population change in *Montastrea annularis* (boulder star coral) and *M. franksi* (boulder star coral) from the Florida Keys, which they attributed to recovery of the corals from the thermal stress of the 1997-1998 ENSO event. The results of our study could suggest that the prevalence of *Symbiodinium* clade A dominance in Millepores is higher before and during a bleaching event and decreases after the event. This change in symbiont population could help the host cope with increased thermal stress during the bleaching event and then help the host start to recover following the bleaching event. Alternatively, the decrease in clade A dominance could also have been a result of a larger sample size in 2017 compared to both 2013 and 2015 collections.

Belize Millepore samples collected from deeper reefs (5-24 m) predominantly contain B as their dominant clade. Generally, temperature and irradiance decrease with depth [50]. Therefore, Millepores that reside on deeper reefs may experience less thermal stress and lower irradiance levels. Deep Belize Millepores examined in this study were 80% *Symbiodinium* clade B dominant, which is in direct contrast to the shallow Belize Millepores that were 72% *Symbiodinium* clade A dominant. *Symbiodinium* clade B dominance in Millepores from deeper reefs may be because these colonies do not require the thermal tolerance conferred by *Symbiodinium* clade A. Finney *et al.* [51] examined the symbiont populations of six orders of Cnidaria from Barbados and Belize, and found that habitat depth influenced the population of symbionts. In their examination of *M. alicornis* and *M. complanata* at shallow depths from Belize, they found both *Symbiodinium* clade A4a and *Symbiodinium* clade B. However, at depths below 5 m with lower ambient light, they no longer found *Symbiodinium* clade A in the symbiont populations of the Millepores. This is a similar trend to the one observed in the current study, and indicates that *Symbiodinium* clade A dominant colonies may be less prevalent on deeper reefs because Millepores may be under less thermal stress at deeper depths.

5. Conclusions

Recent estimates of the impact of climate change predict a 70-90% decline in coral reefs by 2050 [2], which would be devastating for the biodiversity they support and the communities that depend on coral reefs. During the 20th century, SST have risen by 0.74°C and are projected to increase by as much as 4°C during this century [6]. Therefore, understanding the effects of elevated temperatures on *Symbiodinium*-Millepore symbioses is of the utmost importance. Our study demonstrates that Millepores in warmer waters with more irradiance have different symbiont

populations from their counterparts that reside on reefs with cooler temperatures and less irradiance. The presence of background clades indicates that Millepores in both Belize and The Bahamas have the capability to shuffle their symbiont populations to cope with climate change, and that the *Symbiodinium* clade A dominant Belize Millepores may have already shuffled their symbiont populations to deal with the higher thermal stress they experience in the warmer SST of Belize.

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REFERENCES

- [1] Dube, C., Bourmaud, C., Merciere, A., Planes, S., and Boissin, E. (2019). Ecology, Biology and Genetics of *Millepora* Hydrocorals on Coral Reefs. In *Invertebrates- Ecophysiology and Management*, S. Rey, G. Diarte-Plata, and R. Escamilla-Montes, ed.
- [2] Hoegh-Guldberg, O., Kennedy, E., Beyer, H., McClennen, C., and Possingham, H. (2018). Securing a Long-term Future for Coral Reefs. *Trends In Ecology And Evolution* 33, 936-944.
- [3] Sarkis, S., van Beukering, P., McKenzie, E., Brander, L., Hess, S., Bervoets, T., Looijerstijn-van der Putten, L., and Roelfsema, M. (2013). Total Economic value of Bermuda's Coral Reefs: A Summary. In *Coral Reefs Of The United Kingdom Overseas Territories*, C. Sheppard, ed. (Springer Science and Business Media Dordrecht), pp. 201-211.
- [4] Bapna, M., United Nations Development Programme, United Nations Environment Programme, World Bank, and World Resources Institute - July 2008 (2008). *Roots of Resilience: Growing the Wealth of the Poor* (World Resources 2008).
- [5] Eakin, M., Morgan, J., Heron, S., Smith, T., Liu, G., Alvarez-Filip, L., Baca, B., Bartels, E., Bastidas, C., and Bouchon, C. *et al.* (2010). Caribbean Corals in Crisis. *Plos ONE* 5(11), e13968.
- [6] Hoegh-Guldberg, O., Mumby, P., Hooten, A., Steneck, R., Greenfield, P., Gomez, E., Harvell, C., Sale, P., Edwards, A., and Caldira, K. *et al.* (2007). Coral Reefs Under Rapid Climate Change and Ocean Acidification. *Science* 318, 1737-1742.

- [7] Hughes, T., Kerry, J., Alvarez-Noriega, M., Alvarez-Romero, J., Anderson, K., Baird, A., Babcock, R., Beger, M., Bellwood, D., and Berkemans, R. et al. (2017). Global warming and recurrent mass bleaching of corals. *Nature* *543*, 373-377.
- [8] Brown, B. (1997). Coral bleaching: causes and consequences. *Coral Reefs*. *16*, S129-S138.
- [9] Hoegh-Guldberg, O. (1999). Climate change, coral bleaching and the future of the world's coral reefs. *Marine Freshwater Research* *50*, 839-866.
- [10] Fournier, A. (2013). The story of symbiosis with zooxanthellae, or how they enable their host to thrive in a nutrient poor environment. *Biosciences Master Reviews*. 1-8.
- [11] Baker, A. (2003). Flexibility and Specificity in Coral-Algal Symbiosis: Diversity, Ecology, and Biogeography of *Symbiodinium*. *Annual Review Of Ecology Systematics And Evolution*. *34*, 661-689.
- [12] Baker, A., and Romanski, A. (2007). Multiple symbiotic partnerships are common in scleractinian corals, but not in octocorals: Comment on Goulet (2006). *Marine Ecology Progress Series* *335*, 237-242.
- [13] Davy, S., Allemand, D., and Weis, V. (2012). Cell Biology of Cnidarian-Dinoflagellate Symbiosis. *Microbiology And Molecular Biology Reviews* *76*, 229-261.
- [14] Mumby, P., Flower, J., Chollett, I., Box, S., Bozec, Y., Fitzsimmons, C., Forster, J., Gill, D., Griffith-Mumby, R., and Oxenford, H. et al. (2014). Climate Change and its Effects on Caribbean Coral Reefs. In *Towards Reef Resilience and Sustainable Livelihoods: A Handbook For Caribbean Coral Reef Managers*, D. Kleine, ed. (Exeter, Devon, UK: University of Queensland Press), pp. 52-63.
- [15] Wietheger, A., Starzak, D., Gould, K., and Davy, S. (2018). Differential ROS Generation in Response to Stress in *Symbiodinium* spp. *Biological Bulletin*. *234*, 11-21.
- [16] Buddemeier, R., Baker, A., Fautin, D., and Jacobs, R. (2004). The Adaptive Hypothesis of Bleaching. In *Coral Health And Disease*. pp 427-444. (eds) E. Rosenberg and Y Loya. Springer, Berlin, Heidelberg.
- [17] Franklin, D., Hoegh-Guldberg, O., Jones, R., and Berges, J. (2004). Cell death and degeneration in the symbiotic dinoflagellates of the coral *Stylophora pistillata* during bleaching. *Marine Ecology Progress Series* *272*, 117-130.
- [18] Richier, S., Furla, P., Plantivaux, A., Merle, P., and Allemand, D. (2005). Symbiosis-induced adaptation to oxidative stress. *Journal of Experimental Biology* *208*, 277-285.
- [19] Nielsen, D., Petrou, K., and Gates, R. (2018). Coral bleaching from a single cell perspective. *The ISME Journal* *12*, 1558-1567.
- [20] Berkemans, R., and van Oppen, M. (2006). The role of zooxanthellae in the thermal tolerance of corals: a 'nugget of hope' for coral reefs in an era of climate change. *Proceedings of The Royal Society B: Biological Sciences* *273*, 2305-2312.
- [21] Mieog, J., van Oppen, M., Cantin, N., Stam, W., and Olsen, J. (2007). Real-time PCR reveals a high incidence of *Symbiodinium* clade D at low levels in four scleractinian corals across the Great Barrier Reef: implications for symbiont shuffling. *Coral Reefs* *26*, 449-457.
- [22] Coffroth, M., Poland, D., Petrou, E., Brazeau, D., and Holmberg, J. (2010). Environmental Symbiont Acquisition May Not Be the Solution to Warming Seas for Reef-Building Corals. *PlosONE*. *5(10)*, e13258.
- [23] Baker, A., Starger, C., McClanahan, T., and Glynn, P. (2004). Corals' adaptive response to climate change. *Nature* *430* 741.
- [24] Poland, D., and Coffroth, M. (2017). Trans-generational specificity within a cnidarian-algal symbiosis. *Coral Reefs* *36*, 199-129.
- [25] Palumbi, S., Barshis, D., Traylor-Knowles, N., and Bay, R. (2014). Mechanisms of reef coral resistance to future climate change. *Science* *344*, 895-898.
- [26] Gould, W., Diaz, E., Alvarez-Berrios, N., Aponte-Gonzalez, F., Archibald, W., Bowden, J., Carrubba, L., Crespo, W., Fain, S., and Gonzalez, G. et al. (2018). Fourth National Climate Assessment. <https://nca2018.globalchange.gov/chapter/20/>.
- [27] Ruiz-Ramos, D., Weil, E., and Schizas, N. (2014). Morphological and genetic evaluation of the hydrocoral *Millepora* species complex in the Caribbean. *Zoological Studies*. *53*, 4.
- [28] Lewis, J. (2006). Biology and Ecology of the Hydrocoral *Millepora* on Coral Reefs. In *Advances in Marine Biology*, pp1-55. A. Southward, C. Young, and L. Fuiman, ed. (San Diego, CA: Elsevier Ltd.).
- [29] Tepper, C., Squiers, L., Hay, C., Gorbach, D., Friend, D., Black, B., Greenstein, B., and Strychar, K. (2012). Cryptic Species: A Mismatch between Genetics and Morphology in *Millepora*. *Marine Science*. *2(5)* 57-65.
- [30] Schweinsberg, M., Tollrian, R., and Lampert, K. (2016). Inter- and intra-colonial genotypic diversity in hermatypic hydrozoans of the family Milleporidae. *Marine Ecology*. *38*, e12388.
- [31] Takama, O., Fernandez-Silva, I., Lopez, C., and Reimer, J. (2018). Molecular Phylogeny Demonstrates the Need for Taxonomic Reconsideration of Species Diversity of the Hydrocoral Genus *Millepora* (Cnidaria: Hydrozoa) in the Pacific. *Zoological Science*. *35(2)*, 123-133.
- [32] (2020). Water Temperature - San Salvador, Bahamas. <https://www.watertemperature.org/San-Salvador-Bahamas-Island.html>.
- [33] (2020). Sea Temperature - South Water Caye, Belize. <https://seatemperature.info/south-water-caye-water-temperature.html#:~:text=The%20average%20water%20temperature%20in,%20C%20%2F84.2%20C%20B0F>.
- [34] Samayoa, A., Reyes, S., Karim, Y., Roge-Jones, L., Rueth, M., and Tepper, C. (2017). Patterns of Millepore-*Symbiodinium* associations at two Caribbean locations: San Salvador, The Bahamas and South Water Cay, Belize. *1st Joint Symposium of The Natural History and Geology of The Bahamas* 59-70.
- [35] Correa, A., McDonald, M., and Baker, A. (2009). Development of clade-specific *Symbiodinium* primers for quantitative PCR (qPCR) and their application to detecting clade D symbionts in Caribbean corals. *Marine Biology* *156*, 2403-2411.
- [36] Ulstrup, K., and van Oppen, M. (2003). Geographic and Habitat partitioning of genetically distinct zooxanthellae (*Symbiodinium*) in *Acropora* corals on the Great Barrier Reef.

- Molecular Ecology 12. 3477-3484.
- [37] LaJeunesse, T. (2003). Diversity and community structure of symbiotic dinoflagellates from Caribbean coral reefs. *Marine Biology* 141, 387-400.
- [38] (2020). Smithsonian Tropical Monitoring Networks - Real Time Environmental Monitoring Program. http://nmnhmp.riocean.com/arc_vision.php.
- [39] Goulet, T., Cook, C., and Goulet, D. (2005). Effect of short-term exposure to elevated temperatures and light levels on photosynthesis of different host-symbiont combinations in the *Aiptasia pallida/Symbiodinium* symbiosis. *Limnology and Oceanography*. 50(5) 1490-1498.
- [40] Howells, E., Beltran, V., Larsen, N., Bay, L., Willis, B., and van Oppen, M. (2012). Coral thermal tolerance shaped by local adaptation of photosymbionts. *Nature Climate Change* 2, 116-120.
- [41] Buddemeier, R., and Fautin, D. (1993). Coral bleaching as an adaptive mechanism: a testable hypothesis. *BioScience* 43, 320-326.
- [42] Coralreefwatch.noaa.gov. (2021). *Coral Reef Watch Satellite Monitoring and Modeled Outlooks*. <https://coralreefwatch.noaa.gov/product/vs/data/belize.txt>.
- [43] Venn, A., Loram, J., Trapido-Rosenthal, H., Joyce, D., and Douglas, A. (2008). Importance of Time and Place: Patterns in Abundance of *Symbiodinium* Clades A and B in the Tropical Sea Anemone *Condylactis gigantea*. *Biological Bulletin* 215, 243-252.
- [44] Reynolds, J., Bruns, B., Fitt, W., and Schmidt, G. (2008). Enhanced photoprotection pathways in symbiotic dinoflagellates of shallow-water corals and other cnidarians. *PNAS* 105, 13674-12678.
- [45] Kemp, D., Hernandez-Pech, X., Iglesias-Prieto, R., Fitt, W., and Schmidt, G. (2014). Community dynamics and physiology of *Symbiodinium* spp. before, during, and after a coral bleaching event. *Limnology and Oceanography*. 59(3) 788-797.
- [46] De Weerd, W. (1984). Taxonomic characters in Caribbean *Millepora* species (Hydrozoa, Coelenterata). *Bijdragen Tot De Dierkunde* 54, 243-262.
- [47] Thornhill, D., LaJeunesse, T., Kemp, D., Fitt, W., and Schmidt, G. (2005). Multi-year, seasonal genotypic surveys of coral-algal symbioses reveal prevalent stability or post-bleaching reversion. *Marine Biology*. 148, 711-722.
- [48] Fay, S., and Weber, M. (2012). The Occurrence of Mixed Infections of *Symbiodinium* (Dinoflagellata) within Individual Hosts. *Journal Of Phycology* 48, 1306-1316.
- [49] Baker, D., Andras, J., Jordan-Garza, A., and Fogel, M. (2013). Nitrate competition in a coral symbiosis varies with temperature among *Symbiodinium* clades. *The ISME Journal* 7, 1248-1251.
- [50] Forrest, J., Marcucci, E., and Scott, P. (2007). Geothermal Gradients and Subsurface Temperatures in the Northern Gulf of Mexico. *Gulf Coast Association Of Geological Societies Transactions*. 55. 233-248.
- [51] Finney, J., Pettay, D., Sampayo, E., Warner, M., Oxenford, H., and LaJeunesse, T. (2010). The Relative Significance of Host-Habitat, Depth, and Geology on the Ecology, Endemism, and Speciation of Coral Endosymbionts in the Genus *Symbiodinium*. *Microbial Ecology* 60, 250-263.