

# The Ecology of the Saline Lakes in the Semiarid Pampa Central (Argentina): Limnologic Characterization and Zooplankton of Utracán

Echaniz Santiago\*, Cabrera Gabriela, Vignatti Alicia

Departamento de Ciencias Biológicas, Facultad de Ciencias Exactas y Naturales, Universidad Nacional de La Pampa,  
La Pampa, República Argentina

**Abstract** In the semi-arid central Argentina exist temporary saline lakes that have recently begun to be studied. If salinity does not exceed  $30 \text{ g.L}^{-1}$  can be studied using the "alternative states of shallow lakes model". The aim of this study was to determine the ecology of a limnologically unknown lake during a mesosaline period and to verify that due to the salinity, zooplankton richness, density and biomass were similar to other lakes in the region and, due to the fishes' lacking, the zooplankton is consisted of larger species, which contributes to the clear state. The depth and salinity remained stable ( $2 \pm 0.15 \text{ m}$  and  $32.9 \pm 2.7 \text{ g.L}^{-1}$ ). Although high concentrations of nutrients, transparency exceeded 1.1 m and chlorophyll-*a* concentration was around  $1 \text{ mg.m}^{-3}$ . The ratio  $Z_m / Z_{\text{phot}}$  was 0.62 (clear state). Two cladocerans, two copepods and two rotifers were registered and no correlation between richness and environmental parameters was found. The greatest density and biomass was provided by copepods, especially *Boeckella poopoensis* Marsh, 1906. The zooplankton biomass was lower than other Pampean lakes, probably by the limited availability of food derived of the clear state. A particularity of the lake was the reduced influence of environmental factors on zooplankton, probably due to varied in narrow ranges and the species remained within its tolerances ranges.

**Keywords** Saline lakes, Clear state, Zooplankton, *Boeckella poopoensis*

## 1. Introduction

The province of La Pampa is located in the central semiarid of Argentina. It is in a region where the mean annual rainfall show a steep gradient, with 700 mm to the east to less than 300 mm to the west [1], but they are always exceeded by evapotranspiration, which is around 800 to 850 mm per year in the whole region [2].

In this area there are numerous shallow lakes located in arheic basins, primarily fed by rainfall and to a lesser extent by groundwater contributions. Due to the high evaporation of the region, one of the main characteristics of these lakes is the temporality, which makes that containing water for a few weeks or a few years, after which it can dry completely, also for several years. Most of these lakes have concentrations of dissolved solids higher than  $3 \text{ g.L}^{-1}$ , so they can be classified as saline lakes [3-6], but the dynamics given by the filling-drying make their salinity widely vary [7-10].

It is known that salinity is one of the abiotic factors that

mainly influences the biota of these lakes [5, 11, 12], since both zooplankton richness and density are decreasing with increasing the concentration of dissolved solids [3, 13-18]. This phenomenon has been recently verified in La Pampa [5, 6, 12].

While the salinity of these lakes does not exceed about  $30 \text{ g.L}^{-1}$  and preventing the presence of large size cladocerans, the theoretical framework that has proved to be suitable for study is the alternative states of shallow lakes model [19, 20, 21]. This indicates that the presence of planktivorous fishes in an aquatic ecosystem produces an effect of trophic cascade since the predation exerted by fishes on larger size zooplankton species with greater filtration efficiency, leads to increased phytoplankton chlorophyll-*a* concentration and a decrease in water transparency (turbid state) [19, 21-24]. On the other hand, the absence of planktivorous fishes (or low density) favours the development of large size zooplankton cladocerans, especially of the genus *Daphnia*, which due to its high rate of grazing, maintain low concentrations of chlorophyll-*a* and, consequently, the transparency of the water is high (clear state) [19, 21-24]. This model has been tested in some relatively saline lakes of La Pampa, where the presence of the halophilic *Daphnia menucoensis* Paggi, 1996 determines clear states [8, 12, 25].

The limnological characteristics of saline, shallow and

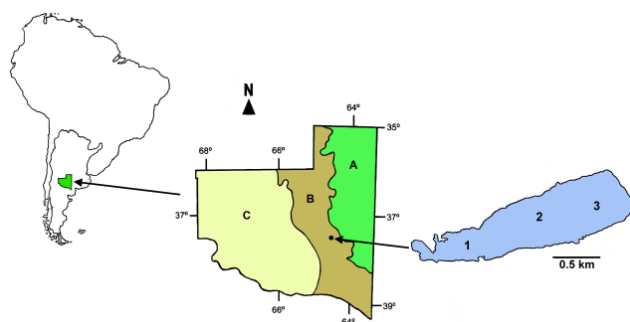
\* Corresponding author:

santiagoechaniz@exactas.unlpam.edu.ar (Echaniz Santiago)

Published online at <http://journal.sapub.org/als>

Copyright © 2015 Scientific & Academic Publishing. All Rights Reserved

temporary lakes elsewhere are known, but in Argentina this type of ecosystems have not received much attention, except for some relatively stable saline lakes in the provinces of Buenos Aires [26-28] and Santa Fe [29]; highland saline lakes in the northwest of the country [30, 31], the Mar Chiquita lake, in Córdoba province [32]. In the central semiarid region of Argentina this kind of lakes is abundant but only recently has begun the study of some of them, pointing out a high variation in their physicochemical and biological characteristics [9, 10]. Because of that, the aim of this study is to describe the variation of the physical, chemical and biological parameters (taxonomic composition, richness, density and zooplankton biomass) and to establish the relationships between them, into a shallow saline lake in the province of La Pampa limnologically unknown, during a period in which the salinity remained within the range mesosaline [3]. In addition to them, another scope of this study is the verification of the following hypotheses: i) due to environmental stress caused by high salinity, zooplankton species richness is reduced, with the presence of halotolerant species; ii) the zooplankton richness, density and biomass are similar to the other lakes of La Pampa having the same range of salinity and iii) due to temporality and high salinity, the lake lacks predatory fishes, so the zooplankton community is mostly composed of large size species, with cladocerans that determine the lake have reduced concentration of phytoplankton chlorophyll-a and consequently, high water transparency (clear state).



**Figure 1.** Geographic location of Utracán Lake. A, B and C: Phytogeographical provinces of the Pampa Plains, Thorny Forest and Monte respectively. 1, 2 and 3: Sampling sites

## 2. Materials and Methods

### 2.1. Study Area

The Utracán Lake (64°36'W, 37°17'S) (Figure 1) is temporary, of an elongate shape, and follows the direction of a depression in a valley of the same name. During the study it had a maximum length and width of 2333 and 649 m, respectively, an area of 96.6 ha and a maximum depth of 2.2 m. It is located in the phytogeographical province of the Thorny Forest [33], surrounded by natural vegetation, with grasslands of *Hyalis argentea* D. Don. Ex Hook. & Arn. and *Stipa brachychaeta* Godron and native forest of *Prosopis*

*caldenia* Burkart and *Geoffroea decorticans* (Gill. Ex Hook. & Arn.) Burkart.

It is largely fed by groundwater contributions and to a lesser extent by rainfall. It is in an arheic basin, so the water losses are mainly due to evaporation. In the basin there is only semi-extensive livestock and there is no current direct influence from urban settlements, since the nearest town is about 15 km away.

During the study, the bed of the lake was almost half-covered with *Ruppia cirrhosa* (Petagna, Grande), a cosmopolitan rhizomatous macrophyte, characteristic of marine coastal or salt ponds (34). It is an herbaceous rooted plant capable of tolerating changes in temperature, salinity and water level fluctuations (35). It had its greatest development in February and March, but then it declined until its disappearance during May. It reappeared again in November and December in strands at the shores. During the study, the lake lacked fishes.

The avifauna was composed mainly of flamingos (*Phoenicopterus chilensis* Molina, 1782), present throughout the year in large flocks, a few white-necked swans (*Coscoroba coscoroba* Molina, 1782) and occasional shorebirds such as *Vanellus chilensis* (Molina, 1782) and *Calidris bairdii* Coves, 1861.

### 2.2. Field and Laboratory Work

Monthly samples of water and zooplankton were collected in the period January- December 2007 at three sampling sites (Figure 1). The water temperature, the concentration of dissolved oxygen (Lutron® OD5510 oximeter), the water transparency (Secchi disk) and the pH (Corning® PS15 pH meter) were recorded at each station, and water samples, which were kept refrigerated until analysis, were collected for physico-chemical determinations.

In addition, a qualitative sample of zooplankton was taken by vertical and horizontal dragging, with a net of 22 cm in diameter at the mouth and mesh opening of 0.04 mm, while quantitative samples were taken with a 10 L Schindler-Patalas plankton trap with a net with a 0.04 mm mesh opening, by means of two vertically aligned extractions, so as to integrate the water column. Samples were anesthetized with CO<sub>2</sub> and kept refrigerated until analysis. The maximal depth of the lake was measured by probing and the length and width with a GPS Garmin® E-Trex Legend.

The dissolved solids concentration (salinity) was determined by the gravimetric method with drying at 104°C of 50 ml of previously filtered water. The chlorophyll-a concentration was determined by extraction with aqueous acetone to 90% and subsequent reading in a spectrophotometer [36, 37], total Kjeldahl nitrogen (TKN) by the Kjeldahl method and total phosphorus (TP) using the method of ascorbic acid, after previous digestion with potassium persulfate. The content of suspended solids was determined with Microclar FFG047WPH fiberglass filters, dried at 103–105°C until constant weight and later calcined at 550°C [38].

Counts of macro and microzooplankton [20] from each sample were carried out under stereoscopic and conventional optical microscopes using Bogorov and Sedgwick-Rafter cameras, respectively. Once analyzed, the samples were fixed with formaline 5% and then deposited in the plankton collection of the Facultad de Ciencias Exactas y Naturales de la Universidad Nacional de La Pampa. To determine the biomass of the zooplankton, a minimum of 30 specimens of all species were measured with a Leitz ocular micrometer and formulas that relate to the total length with the dry weight of the specimens were used [39-43].

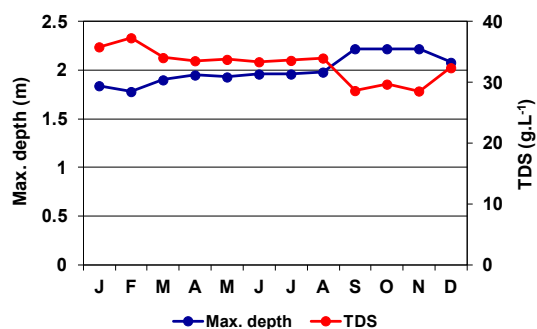
We used the classification of continental waters based on salinity proposed by Hammer [3]. In order to examine the relationships between environmental factors and attributes of zooplankton, nonparametric Spearman correlation coefficients ( $r_s$ ) were calculated [44-46].

In order to characterize the lake as a function of its transparency, we calculated the relationship between the mean depth and that of the photic zone:  $Z_m/Z_{phot}$  [47]. If the calculated value is less than 1, it is considered as a clear lake, and if this ratio is greater than 1, it is a turbid one. For the calculation of the depth of the photic zone, we multiplied the reading of the Secchi disk by a factor of 3 [20, 48]. We used Past [49] and Infostat [50] software.

### 3. Results

#### 3.1. Environmental Parameters

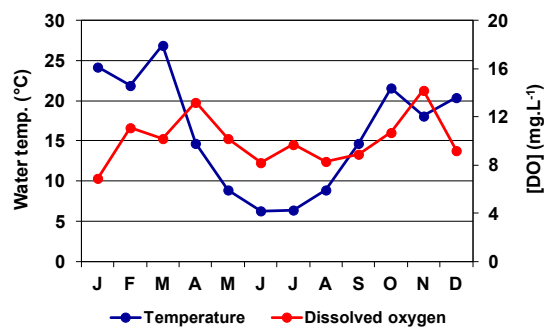
The depth of the lake was slightly lower when the study began, at around 1.8 m (January and February). It rose in March to about 1.95 m and it remained relatively stable until September, when increased and reached the maximum depth recorded (2.22 m) (Figure 2). The mean dissolved solids concentration was  $32.9 \text{ g.L}^{-1} (\pm 2.7)$ ; this was also quite stable and followed a reverse pattern with depth, so that a high correlation was found between these two parameters ( $r_s = -0.91$ ;  $p = 0.0000$ ) (Figure 2).



**Figure 2.** Monthly variation of the maximum depth and total dissolved solids in Utracán Lake during 2007

The water temperature followed a well-defined seasonal pattern. In the winter arrived to around  $6^\circ\text{C}$  and in summer reached almost  $27^\circ\text{C}$  (Figure 3). The concentration of dissolved oxygen, although it varied throughout the study period, did not show a seasonal pattern (Figure 3). It had a

minimum in January ( $6.9 \text{ mg.L}^{-1}$ ) and reached the maximum in November ( $14.2 \text{ mg.L}^{-1}$ ). No significant correlation was found between the two parameters.

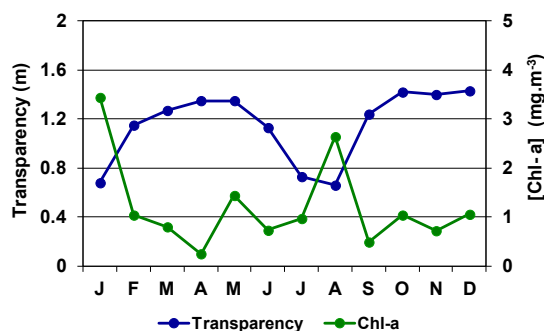


**Figure 3.** Monthly variation of the water temperature and dissolved oxygen in Utracán Lake during 2007

The ionic composition of the water showed a strong predominance of  $\text{Na}^+$ , presenting a mean concentration of  $9087.75 \text{ mg.L}^{-1} (\pm 1560)$ , it represented 92% of the cations, while the divalents  $\text{Ca}^{++}$  and  $\text{Mg}^{++}$  showed reduced concentrations ( $345.9 \text{ mg.L}^{-1} \pm 63.9$  y  $126.1 \text{ mg.L}^{-1} \pm 23.5$ , respectively), which maintained a relatively low hardness of the water.  $\text{Cl}^-$ , with a mean concentration of  $7853 \text{ mg.L}^{-1} (\pm 3738.2)$ , represented almost 44% of the anions while the rest anions showed lower and more relatively similar concentrations. The pH was relatively stable throughout the year, with a mean of  $9.58 (\pm 0.13)$ .

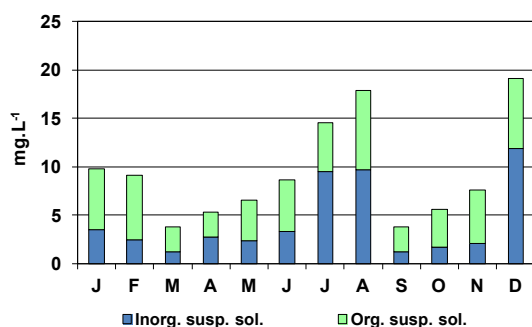
The mean water transparency was higher than 1.1 m (Table 1). The lowest value (0.7 m) was recorded in August, while the Secchi disk reading values showed higher values for nine months, exceeding 1 meter, with a maximum in December (1.43 m) (Figure 4). The calculation of the  $Z_m/Z_{phot}$  ratio shows a value of 0.62.

The mean annual concentration of chlorophyll-*a* was very low, around  $1 \text{ mg.m}^{-3}$  (Table 1). It did not show a clear seasonal pattern, since although the minimum concentration ( $0.25 \text{ mg.m}^{-3}$ ) was recorded in early fall, in April, two peaks were recorded, one in summer (January:  $3.45 \text{ mg.m}^{-3}$ ) and the other in winter (August:  $2.64 \text{ mg.m}^{-3}$ ) (Figure 4). Although some correspondence was observed, no correlation was found between water transparency and phytoplankton chlorophyll-*a* concentration.



**Figure 4.** Monthly variation of the water transparency and chlorophyll-*a* concentration in Utracán Lake during 2007

The highest concentration of total suspended solids, close to  $19 \text{ mg.L}^{-1}$ , was registered in December (Figure 5). With the exception of three months in which the inorganic fraction was relatively high, the suspended solids of organic origin always slightly dominated (Figure 5). No correlation was found between water transparency and total suspended solids or both fractions separately.



**Figure 5.** Monthly variation of the suspended solids concentration in Utracán Lake during 2007

The mean concentrations of nutrients in the water were high, since they exceeded 7 and  $13 \text{ mg.L}^{-1}$  in the case of TP and TKN, respectively (Table 1). In both cases their values were higher in winter, since in July the TP peaked and reached  $17.5 \text{ mg.L}^{-1}$  while TKN reached  $24.4 \text{ mg.L}^{-1}$ . However, only a negative correlation was found with the temperature in the second case ( $r_s = -0.59$ ;  $p = 0.0417$ ).

**Table 1.** Mean values of the principal environmental parameters determined in Utracán Lake during 2007 (the number in brackets indicates the standard deviation)

Total dissolved solids ( $\text{g.L}^{-1}$ )	32.90 ( $\pm 2.69$ )
Water temperature ( $^{\circ}\text{C}$ )	16.08 ( $\pm 7.17$ )
Dissolved oxygen ( $\text{mg.L}^{-1}$ )	10.07 ( $\pm 2$ )
Transparency (m)	1.15 ( $\pm 0.29$ )
Chlorophyll- <i>a</i> ( $\text{mg.m}^{-3}$ )	1.22 ( $\pm 0.92$ )
pH	9.58 ( $\pm 0.13$ )
Total suspended solids ( $\text{mg.L}^{-1}$ )	9.30 ( $\pm 5.22$ )
Inorganic suspended solids ( $\text{mg.L}^{-1}$ )	4.30 ( $\pm 3.77$ )
Organic suspended solids ( $\text{mg.L}^{-1}$ )	4.99 ( $\pm 1.89$ )
Total phosphorus ( $\text{mg.L}^{-1}$ )	7.21 ( $\pm 3.75$ )
Total Kjeldahl nitrogen ( $\text{mg.L}^{-1}$ )	13.03 ( $\pm 4.87$ )

### 3.2. Zooplankton

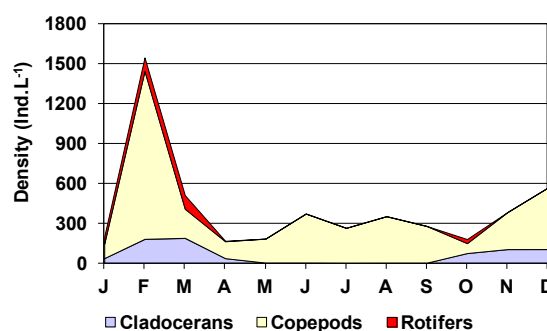
Six *taxa* were recorded: two cladocerans, two copepods and two rotifers (Table 2). The greatest richness was recorded in mid-fall (May) when all the six species were found. The lowest richness was recorded during the coldest months (June, July and August) when only three species of crustaceans and no rotifers were recorded. However, no correlation was found between total richness and the water temperature or salinity. Considering these separately, a significant correlation was found between the richness of rotifers and water temperature ( $r_s = 0.59$ ;  $p = 0.0419$ ) because

they were only recorded during the months with higher temperatures.

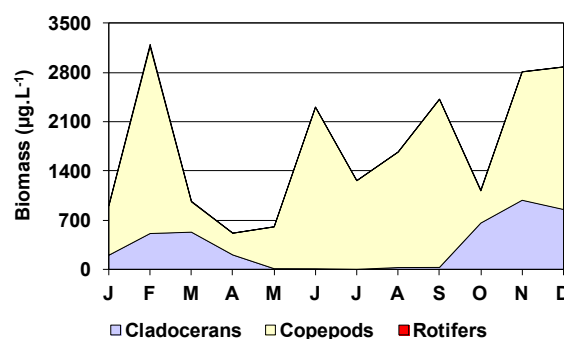
Among crustaceans, the calanoid *Boeckella poopoensis* Marsh, 1906 was recorded on all sampling occasions, followed by *Moina eugeniae* Olivier, 1954 and *Cletocamptus deitersi* (Richard, 1897), found on 10 samples. In terms of rotifers, *Hexarthra fennica* (Levander, 1892) was found in eight samples but *Brachionus plicatilis* Müller, 1786 only in five (Table 2).

The total zooplankton density and biomass were higher in summer, as in February they reached  $1548.7 \text{ ind.L}^{-1}$  and  $3208.8 \text{ } \mu\text{g.L}^{-1}$ , respectively, and lower during the fall, when they went down to a minimum of  $166 \text{ ind.L}^{-1}$  and  $517 \text{ } \mu\text{g.L}^{-1}$ , respectively (April). The greatest density and biomass values were provided by calanoid copepods, although their behavior varied throughout the study. During the summer a peak a high density was recorded and nauplius larvae predominated, accounting for 80% of the density (Figs. 6 and 7). At other times the density was lower, but high biomasses were reached due to the dominance of adults and copepodites (Figs. 6 and 7). Cladocerans were recorded during the warmer months, when they contributed a relatively high biomass, so what happened during the spring, when they provided 58% of the total biomass in October (Figs. 6 and 7).

Rotifers were always recorded in relatively low densities and their contribution to total biomass was always less than 1% (Figs. 6 and 7). No significant correlations were found between the total zooplankton density and biomass and different environmental parameters analyzed.



**Figure 6.** Variation in the composition of the zooplankton density of the Utracán Lake during 2007

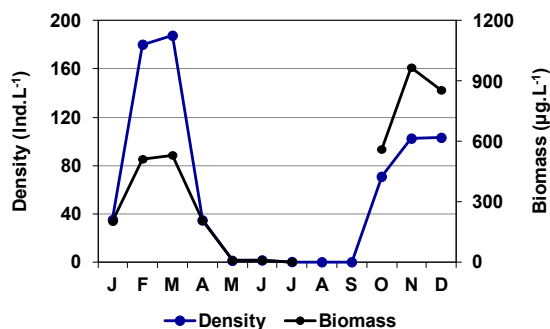


**Figure 7.** Variation in the composition of the zooplankton biomass of the Utracán Lake during 2007



### 3.3. Density, Biomass and Size of the Main Species

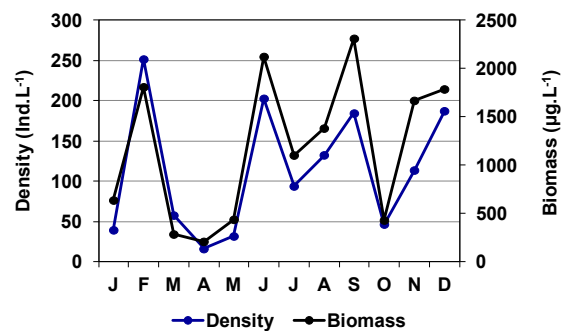
Among cladocerans, *M. eugeniae* was recorded in ten samples. It presented a maximum density of 187.8 ind.L<sup>-1</sup> in late summer (March), but as the specimens were of a small size (mean = 784.64  $\mu\text{m} \pm 173.48$ ) its biomass was not so high (Figure 8). The density and the biomass of this cladoceran decreased in late fall and early winter and it was not recorded in August and September. It reappeared in mid-spring, in October, after which the largest biomass reached a peak of 967.6  $\mu\text{g.L}^{-1}$  in November, although as larger specimens were found (mean = 1107.33  $\mu\text{m} \pm 247.89$ ), the density was not as high as in summer (Figure 8). The mean size of *M. eugeniae* (Table 2) showed a significant correlation with salinity ( $r_s = -0.74$ ;  $p = 0.0365$ ). *Daphnia menucoensis* Paggi, 1996 was recorded only five times and its density was reduced (Table 2). Except for an occasional record in May, it was only found from early spring. It reached the greatest density and biomass in October, with 2.33 ind.L<sup>-1</sup> and 100.2  $\mu\text{g.L}^{-1}$ , respectively. Correlations between the mean size of this species (Table 2) and the concentration of total suspended solids were found, particularly with those of organic origin ( $r_s = -0.90$ ;  $p = 0.0374$ ).



**Figure 8.** Changes in the density and biomass of *Moina eugeniae* in the Utracán Lake during 2007

*B. poopoensis* was present during the whole year, and

unlike the previous cases, presented greater fluctuations throughout the annual cycle without exhibiting a seasonal pattern. The maximum density was recorded in February (251.7 ind.L<sup>-1</sup>), but as the specimens were smaller (mean = 1021.77  $\mu\text{m} \pm 326.15$ ), its biomass was not so high. Conversely, during the winter and early spring there were found larger specimens (mean = 1184  $\mu\text{m} \pm 287.56$  in July and 1199  $\mu\text{m} \pm 247.27$  in September), so higher biomasses were recorded, with a peak of 2312  $\mu\text{g.L}^{-1}$  in September (Figure 9). No significant correlations between the size of this species and the environmental variables considered were found.



**Figure 9.** Changes in the density and biomass of *Boeckella poopoensis* in the Utracán Lake during 2007

Among rotifers, *B. plicatilis* was only present for the first five months and reached a relatively high density and biomass in February (103.3 ind.L<sup>-1</sup> and 25.8  $\mu\text{g.L}^{-1}$ , respectively). On other sampling occasions, it never exceeded 2 ind.L<sup>-1</sup> or 0.5  $\mu\text{g.L}^{-1}$ . *H. fennica*, present for eight months, was absent during the winter. Its highest density was recorded in March at 101.7 ind.L<sup>-1</sup> but it only reached a biomass of 7.2  $\mu\text{g.L}^{-1}$  because the mean size was small (mean = 159.6  $\mu\text{m} \pm 23.88$ ); however, in October it reached 10.54  $\mu\text{g.L}^{-1}$  when a lower density was recorded (10.5 ind.L<sup>-1</sup>) but the mean size of the specimens was higher (mean = 264.95  $\mu\text{m} \pm 31.23$ ).

**Table 2.** Species registered and frequency of their occurrence in the samples (%) during the studied annual cycle in the Utracán Lake

	Frequency (%)		Density (Ind.L <sup>-1</sup> )	Biomass (µg.L <sup>-1</sup> )	Body length µm
<i>Daphnia menucoensis</i> Paggi, 1996	41.7	Mean	0.36	14.76	1217.7
		Min. – max.	0–2.3	0–100.2	715–2320
<i>Moina eugeniae</i> Olivier, 1954	83.3	Mean	59.76	321.16	953.03
		Min. – max.	0–187.8	0–967.6	572–1601.6
<i>Boeckella poopoensis</i> Marsh, 1906	100	Mean	113.13	1181.65	1146.60
		Min. – max.	16–251.7	2082311.8	543.4–1930.5
<i>Cletocamptus deitersi</i> (Richard, 1897)	83.3	Mean	1.94	11.48	802.87
		Min. – max.	0–5.8	0–38.8	314.4–1072.5
<i>Brachionus plicatilis</i> Müller, 1786	41.7	Mean	9.96	2.51	276
		Min. – max.	0–103.3	0–25.8	214.5–357.5
<i>Hexarthra fennica</i> (Levander, 1892)	66.7	Mean	14.7	1.87	198.9
		Min. – max.	0–101.7	0–10.5	100.1–314.6

## 4. Discussion

Despite the temporary nature of Utrac  n – it was revealed both before and after 2007 that the lake level showed significant declines (Echaniz & Vignatti, pers. obs.) – during the studied period its depth and its salinity remained relatively stable. The oscillations barely reached 0.4 m and 9 g.L<sup>-1</sup>, which kept it in the mesosaline range [3], while during the same period other Pampean lakes recorded declines in water level above 25% and salinity increases reached double the initial values [10].

The chemical composition of the water, with the predominance of Na<sup>+</sup> between cations and the Cl<sup>-</sup> between anions, is a characteristic similar to most of the shallow lakes in the province and indicates that the mechanisms predominate in controlling water chemistry involving evaporation and crystallization. This is a typical situation in arid or semiarid regions such as the central region of La Pampa, where evapotranspiration rates are higher than rainfall [20, 51, 52].

The low value of the  $Z_m/Z_{phot}$  ratio found allows Utrac  n to be characterized as a clear lake and the phytoplankton chlorophyll-*a* concentrations and suspended solids were reduced, which differs from other shallow lakes of La Pampa, which are strongly turbid. In many, the reduced transparency is due to high concentrations of phytoplankton chlorophyll, a situation promoted by the absence of large-sized zooplankton because of predation by zooplanktivorous fishes [25, 53, 54], and in others lakes to high concentrations of inorganic suspended solids in the water column [10, 55, 56]. The high transparency of the water of Utrac  n may be the result of, on the one hand, the presence of cladocerans of a relatively large size and therefore efficient filtering, due to the absence of predation by fishes, which cannot colonize this ecosystem because of its saline and temporary nature. On the other hand, the transparency allowed light to reach the bottom, which led to the development of *R. cirrhosa*. The fact of having a large proportion of the bed covered by the macrophyte has contributed to the fact that the sediments were not resuspended by wind. However, the minimum transparency, produced by inorganic suspended solids, was recorded in August, the month with a historical record of strongest winds in the region [57].

The concentrations of nutrients measured were so high that it allowed Utrac  n to be categorized as a hypertrophic lake [58]; however, they were similar to other environments in La Pampa [6, 59, 60, 61]. This could well attributed in part to the drag, especially during storms, of the excreta of animals that graze in the basin [56, 62, 63, 64], and partly to the resolubilization of nutrients from the sediments [65, 66]. The latter is a particularly important phenomenon in shallow lakes [67-69], and in Utrac  n it coincided with the highest concentrations of nutrients in the water being recorded during July and August, the months with the strongest winds [57].

In La Pampa it has been found that subsaline lakes not exceeding 2 g.L<sup>-1</sup> may have more than 20 taxa in their

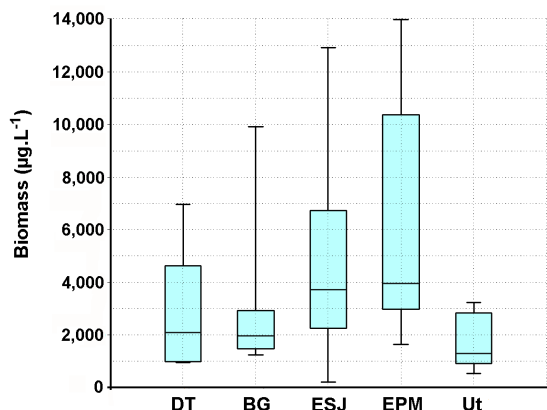
zooplankton [4, 54, 59, 61], while in hypo-mesosaline lakes, where salinity has a major influence in the structuring of their community [3, 11, 18], the richness is lower [60]. The latter situation was verified in Utrac  n, where the diversity of zooplankton was reduced and the association is characterized particularly by halophilic native species, such as *D. menucoensis*, *M. eugeniae* and *B. poopoensis* [6, 59, 60, 70]. The two first species are very common in saline ecosystems of northern Patagonia and the central region of the country [10, 12, 59, 60, 71], while *B. poopoensis* shows a much wider geographical distribution, from the north of the Patagonian plateau to the south of Peru [30, 72, 73]. Among rotifers, the presence of eurihaline species such as *B. plicatilis* and *H. jennica* is a commonly characteristic with other similar water bodies in the province [12, 59, 60, 74], although there are species that are not typical of the region due their cosmopolitan distribution [75].

On the other hand, the prevalence of crustaceans, especially with the high density of *B. poopoensis*, and the limited participation of rotifers are a characteristic point which distinguishes Utrac  n from other saline lakes without fishes, where the highest density was provided by species of the genus *Brachionus*, particularly *B. plicatilis*, *B. dimidiatus* and *B. angularis* [8, 9, 10].

The zooplankton biomass recorded in Utrac  n was much lower than that recorded currently in other shallow lakes in La Pampa (Figure 10). It was almost the half that recorded in Don Tom  s, a subsaline lake adjacent to the city of Santa Rosa, or 40% lower than that found in the hyposaline Bajo de Giuliani. This occurred despite the fact that in both lakes fish fauna with a predominance of zooplanktivorous fishes was recorded, which suggests that most of the zooplankton species were smaller but they reached very high densities, with a predominance of cyclopoid copepods and rotifers [54, 76]. The Utrac  n biomass was also between 2.7 and 3.7 times less than that of the lakes located in San Jose and Pey-Ma, respectively [9, 10]. This could be attributed to the fact that, although they are also mesosaline lakes and lack fishes like Utrac  n, during the annual cycles studied phytoplankton chlorophyll-*a* concentrations were between 20 and 30 times higher, and that of suspended solids 3 to 7 times higher, which could suggest a much larger supply of food for zooplankton. In Utrac  n, it is likely that the shadowing effect caused by the presence of *R. cirrhosa* may have limited the development of phytoplankton, and thus the amount of food available, which may have prevented as high a development of biomass as in the San Jose and Pey-Ma lakes. However, one point in common among these ecosystems is the predominance of calanoid copepods, especially *B. poopoensis*, since in all three lakes this was the species that mostly contributed to the biomass.

Although the zooplankton diversity was low and characterized by the typical halotolerant species of the region, a particular point was recorded in Utrac  n during this study: the relatively small influence of environmental factors on the density and biomass of zooplankton. This may be because, except for the water temperature, the most relevant

environmental parameters, such as salinity and chlorophyll-a concentration, varied in relatively narrow ranges. The species recorded in Utracán were probably within their range of tolerance, so they would not have shown signs of environmental stress.



**Figure 10.** Comparisons of the mean biomass of five shallow lakes in La Pampa. DT: Don Tomás (subsaline with fish) [61]. BG: Bajo de Giuliani (hyposaline with fish [53]. ESJ: Estancia San José (mesosaline without fish) [10]. EPM: Estancia Pey-Ma (mesosaline without fish) [9]. Ut: Utracán (mesosaline without fish)

## 5. Conclusions

In La Pampa province (central Argentina) exist many temporary saline lakes whose ecology is poorly known, among them is Utracán Lake. The study of the lake during a mesosaline period showed that the zooplankton richness was very low, with the predominance of the calanoid *Boeckella poopoensis* Marsh, 1906. Two particularities of Utracán were the reduced influence of environmental factors on zooplankton, probably because they remained relatively stable and the species were always within their tolerances ranges and that zooplankton biomass was lower than other Pampean lakes, probably by the scarce availability of food.

## REFERENCES

- [1] Casagrande, G., G. Vergara and Y. Bellini, 2006. Cartas agroclimáticas actuales de temperaturas, heladas y lluvias de la provincia de La Pampa (Argentina). *Revista de la Facultad de Agronomía, UNLPam*, 17(1/2): 15 – 22.
- [2] Ponce de León, E., 1998. *Evapotranspiración*. Pag. 31-42. En Fundación Chadileuvú (eds.). *El agua en La Pampa*. Fondo Editorial Pampeano, Santa Rosa.
- [3] Hammer, U. T., 1986. *Saline Lake Ecosystems of the World*. Monographiae Biologicae 59. Dr. W. Junk Publishers, Dordrecht.
- [4] Echaniz, S. A. and A. M. Vignatti, 2010. Diversity and changes in the horizontal distribution of crustaceans and rotifers in an episodic wetland of the central region of Argentina. *Biota Neotropica*, 10(3): 133-141.
- [5] Echaniz, S. A. and A. M. Vignatti, 2011. Seasonal variation and influence of turbidity and salinity on the zooplankton of a saline lake in central Argentina. *Latin American Journal of Aquatic Research*, 39(2): 306-315.
- [6] Vignatti, A. M., G. C. Cabrera and S. A. Echaniz, 2012. Changes in the zooplankton and limnological variables of a temporary hypo-mesosaline wetland of the central region of Argentina during the drying. *Pan American Journal of Aquatic Sciences*, 7(2): 93-106.
- [7] Vignatti, A. M., R. Festa, G. C. Cabrera and S. A. Echaniz, 2012. Comparación luego de una década de parámetros limnológicos, riqueza y abundancia del zooplancton de un lago somero salino de La Pampa. *BioScriba*, 5(1): 23-35.
- [8] Vignatti, A. M., J. C. Paggi, S. A. Echaniz and G. C. Cabrera, 2012. Zooplankton diversity and relationship with environmental changes after the filling of a temporary saline lakes in the semi-arid region of La Pampa (Argentina). *Latin American Journal of Aquatic Research*, 40(4): 1005-1016.
- [9] Echaniz, S. A., G. C. Cabrera, C. Rodríguez and A. M. Vignatti, 2013. Do temporary lakes vary from year to year? A comparison of limnological parameters and zooplankton from two consecutive annual cycles in an Argentine temporary saline lake. *International Journal of Aquatic Sciences*, 4: 44-61.
- [10] Echaniz, S. A., G. C. Cabrera, P. L. Aliaga and A. M. Vignatti, 2013. Variation in zooplankton and limnological parameters in a saline lake of La Pampa, Central Argentina, during an annual cycle. *International Journal of Ecosystem*, 3(4): 72-81.
- [11] Herbst, D., 2001. Gradients of salinity stress, environmental stability and water chemistry as a templet for defining habitat types and physiological strategies in inland salt waters. *Hydrobiologia*, 466: 209-219.
- [12] Echaniz, S. A., A. M. Vignatti, S. B. José de Paggi, J. C. Paggi and A. Pilati, 2006. Zooplankton seasonal abundance of South American saline shallow lakes. *International Review of Hydrobiology*, 91(1): 86-100.
- [13] Green, J., 1993. Zooplankton associations in East African Lakes spanning a wide salinity range. *Hydrobiologia*, 267: 249-256.
- [14] Greenwald, G. M. and S. H. Hurlbert, 1993. Microcosm analysis of salinity effects on coastal lagoons plankton assemblages. *Hydrobiologia*, 267: 307-335.
- [15] Williams, W. D., 1998. Salinity as a determinant of the structure of biological communities in salt lakes. *Hydrobiologia*, 381: 191-201.
- [16] Hall, C. and C. Burns, 2003. Responses of crustacean zooplankton to seasonal and tidal salinity changes in the coastal Lake Waiholo, New Zealand. *New Zealand Journal of Marine and Freshwater Research*, 37: 31-43.
- [17] Derry, A. M., E. E. Prepas and P. D. N. Hebert, 2003. Comparison of zooplankton communities in saline lakewater with variable anion composition. *Hydrobiologia*, 505:199-215.
- [18] Ivanova, M. B. and T. I. Kazantseva, 2006. Effect of Water pH and Total Dissolved Solids on the Species Diversity of

Pelagic Zooplankton in Lakes: A Statistical Analysis. *Russian Journal of Aquatic Ecology*, 37(4): 264-270.

- [19] Scheffer, M., 1998. *Ecology of Shallow Lakes*. Chapman & Hall, London.
- [20] Kalff, J., 2002. *Limnology. Inland Water System*. Prentice Hall.
- [21] Scheffer, M., and E. Jeppesen, 2007. Regime shifts in shallow lakes. *Ecosystems*, 10: 1-3.
- [22] Scheffer, M., S. H. Hosper, M. L. Meijer, B. Moss and E. Jeppesen, 1993. Alternative equilibria in shallow lakes. *Trends in Ecology and Evolution*, 8: 275- 279.
- [23] Jeppesen, E., M. Søndergaard, A. Pedersen, K. Jürgens, A. Strzelczak, T. Lauridsen, and L. Johansson, 2007. Salinity Induced Regime Shift in Shallow Brackish Lagoons. *Ecosystems*, 10: 47-57.
- [24] Jeppesen, E., M. Søndergaard, M. Meerhoff, T. L. Lauridsen and J. P. Jensen, 2007. Shallow lake restoration by nutrient loading reduction- some recent findings and challenges ahead. *Hydrobiologia*, 584: 239-252.
- [25] Echaniz, S. A., A. M. Vignatti, S. B. José de Paggi and G. C. Cabrera, 2010. El modelo de estados alternativos de lagos someros en La Pampa: comparación de Bajo de Giuliani y El Carancho. *Libro de Trabajos del 3° Congreso Pampeano del Agua*. 45-53.
- [26] Olivier, S. R., 1955. Contribution to the limnological knowledge of the Salada Grande lagoon. *Proceedings International Association of Limnology*, 12: 302-308.
- [27] Ringuelet, R. A., 1968. Tipología de las lagunas de la provincia de Buenos Aires. La limnología regional y los tipos lagunares. *Physis*, 28(76): 65-76.
- [28] Ringuelet, R. A., 1972. Ecología y Biocenología del habitat lagunar o lago de tercer orden de la región neotrópica templada (Pampasia Sudoriental de la Argentina). *Physis*, 31(82): 55-76.
- [29] José de Paggi, S. B. and J. C. Paggi, 1998. Zooplancton de ambientes acuáticos con diferentes estados tróficos y salinidad. *Neotropica*, 44(111-112): 95-106.
- [30] Locascio de Mitrovich, C., A. Villagra de Gamundi, J. Juárez and M. Ceraolo, 2005. Características limnológicas y zooplancton de cinco lagunas de la Puna – Argentina”, *Ecología en Bolivia*, 40(1): 10-24.
- [31] Villagra de Gamundi, A., C. Locascio de Mitrovich, J. Juárez and G. Ferrer, 2008. Consideraciones sobre el zooplancton de las lagunas de Yala (Jujuy, Argentina). *Ecología en Bolivia*, 43(2): 1-16.
- [32] Bucher, E. H., (Ed.), 2006. *Bañados del río Dulce y laguna Mar Chiquita (Córdoba, Argentina)*. Academia Nacional de Ciencias (Córdoba, Argentina), 215 pp.
- [33] Cabrera, A., 1976. *Regiones fitogeográficas argentinas. Fascículo 1, Enciclopedia Argentina de agricultura y jardinería*. Ed. Acme. Buenos Aires.
- [34] Obrador, B., J. Pretus and M. Menéndez, 2007. Spatial distribution and biomass of aquatic rooted macrophytes and their relevance in the metabolism of a Mediterranean coastal lagoon. *Scientia Marina*, 71(1): 57-64.
- [35] Menéndez, M., D. Carlucci, M. Pinna, F. Comin and A. Basset, 2003. Effect of nutrients on decomposition of *Ruppia cirrhosa* in a shallow coastal lagoon. *Hydrobiologia*, 506/509: 729-735.
- [36] APHA, 1992. *Standard Methods for the Examination of Water and Wastewater*. 18th edition. American Public Health Association (APHA), American Water Works Association (AWWA) and Water Pollution Control Federation (WPCF), Washington, DC.
- [37] Arar, E. J., 1997. *In Vitro Determination of Chlorophylls a, b, c + c and Pheopigments in Marine and Freshwater Algae by Visible Spectrophotometry. Method 446.0*. U.S. Environmental Protection Agency.
- [38] EPA, 1993. *ESS Method 340.2: Total Suspended Solids, Mass Balance (Dried at 103- 105°C) Volatile Suspended Solids (Ignited at 550°C)*. Environmental Protection Agency (EPA). <http://www.epa.gov/glnpo/lmmb/methods/methd340.pdf>.
- [39] Dumont, H. J., I. van de Velde and S. Dumont, 1975. The dry weight estimate of biomass in a selection of Cladocera, Copepoda and Rotifera from the plankton, periphyton and benthos of continental waters. *Oecologia*, 19: 75-97.
- [40] Ruttner-Kolisko, A., 1974. *Plankton rotifers; Biology and taxonomy*. Die Binnengewässer 26 (1), Stuttgart.
- [41] Rosen, R. A., 1981. Length - dry weight relationships of some freshwaters zooplankton. *Journal of Freshwater Ecology*, 1: 225-229.
- [42] McCauley, E., 1984. *The estimation of the abundance and biomass of zooplankton in samples*, pp 228-265 In: Downing, J. A. & F. H. Rigler (eds.). *A manual on methods for the assessment of secondary productivity in freshwaters*. 2ª ed. Blackwell Scientific. Publ. Oxford.
- [43] Culver, D. A., M. Boucherle, D. J. Bean and J. W. Fletcher, 1985. Biomass of freshwater crustacean zooplankton from length-weight regressions. *Canadian Journal Fisheries and Aquatic Sciences*, 42(8): 1380-1390.
- [44] Sokal, R. and F. Rohlf, 1995. *Biometría. Principios y métodos estadísticos en la investigación biológica*. Ed. Blume, Barcelona.
- [45] Zar, J. H., 1996. *Biostatistical analysis*. 3º Ed. Prentice Hall, New Jersey.
- [46] Pereyra, A., N. Abiati and E. Fernández, 2004. *Manual de estadística para proyectos de investigación*. Ed. Fac.de Cs Agrarias, Univ. Nac. de Lomas de Zamora.
- [47] Quirós, R., A. Rennella, M. Boveri, J. Rosso and A. Sosnovsky, 2002. Factores que afectan la estructura y el funcionamiento de las lagunas pampeanas. *Ecología Austral*, 12: 175-185.
- [48] Cole, G.A., 1988. *Manual de limnología*. Ed. Hemisferio Sur, Bs. As.
- [49] Hammer, Ø., D. Harper, and P. Ryan, 2001. PAST: Paleontological Statistics Software Package for Education and Data Analysis. *Palaeontologia Electronica*, 4(1): 1-9.
- [50] Di Rienzo, J.A., F. Casanoves, M. G. Balzarini, L. González, M. C. Tablada and W. Robledo, 2010. *InfoStat (versión 2010)*. Grupo InfoStat, FCA, Universidad Nacional de Córdoba, Argentina.



- [51] Wetzel, R., 2001. *Limnology. Lake and river ecosystems*. Academic Press, Elsevier, San Diego.
- [52] Gibbs, R., 1970. Mechanisms controlling world water chemistry. *Science*, 170: 1088-1090.
- [53] Echaniz, S. A., A. M. Vignatti and G. C. Cabrera, 2009. Características limnológicas de una laguna turbia orgánica de la provincia de La Pampa y variación estacional del zooplancton. *Biología Acuática*, 26: 71-82.
- [54] Echaniz, S. A., A. M. Vignatti, G. C. Cabrera and S. B. José de Paggi, 2012. Zooplankton richness, abundance and biomass of two hypertrophic shallow lakes with different salinity. *Biota Neotropica*, 12(2): 37-44.
- [55] Vignatti, A. M., S. A. Echaniz and G. C. Cabrera, 2012. Changes in the zooplankton and limnological variables of a temporary hypo-mesosaline wetland of the central region of Argentina during its drying. *Pan-American Journal of Aquatic Sciences*, 7(2): 93-106.
- [56] Echaniz, S. A. and A. M. Vignatti, 2013. Trophic status of shallow lakes of La Pampa (Argentina) and its relation with the land use in the basin and nutrient internal load. *Journal of Environmental Protection (Special Issue on Eutrophication)*, 4(11A): 51-60.
- [57] Cano, E., (coord.), 1980. *Inventario Integrado de los Recursos Naturales de la provincia de La Pampa*. Ed. Instituto Nacional de Tecnología Agropecuaria (INTA), Provincia de La Pampa y Universidad Nacional de La Pampa, Buenos Aires.
- [58] OECD (Organization for Economic Cooperation and Development), 1982. *Eutrophication of waters. Monitoring, Assessment and Control*. Final report, Paris.
- [59] Echaniz, S. A., 2010. *Composición y abundancia del zooplancton en lagunas de diferente composición iónica de la provincia de La Pampa*. Tesis Doctoral. Universidad de Río Cuarto, Facultad de Ciencias Exactas, Físico-Químicas y Naturales.
- [60] Vignatti, A. M., 2011. *Biomasa del zooplancton en lagunas salinas y su relación con la concentración de sales en ausencia de peces*. Tesis Doctoral. Universidad Nacional de Río Cuarto Facultad de Ciencias Exactas, Físico-Químicas y Naturales.
- [61] Echaniz, S. A., A. M. Vignatti and P. C. Bunino, 2008. El zooplancton de un lago somero hipereutrófico de la región central de Argentina. Cambios después de una década. *Biota Neotropica*, 8(4): 63-71.
- [62] Carpenter, S., N. Caraco, D. Correll, R. Howarth, A. Sharpley and V. Smith, 1998. Nonpoint pollution of surface waters with phosphorus and nitrogen. *Ecological Applications*, 8(3): 559-568.
- [63] Bennett, E. M., T. Reed-Andersen, J. Houser, J. Gabriel and S. Carpenter, 1999. A phosphorus budget for the lake Mendota watershed. *Ecosystems*, 2: 69-75.
- [64] Bremigan, M., P. Soranno, M. González, D. Bunnell, K. Arend, W. Renwick, R. Stein and M. Vanni, 2008. Hydrogeomorphic features mediate the effects of land use/cover on reservoir productivity and food webs. *Limnology and Oceanography*, 53(4): 1420-1433.
- [65] Havens, K., K-R Jin, N. Iricanin and R. James, 2007. Phosphorus dynamics at multiple time scales in the pelagic zone of a large shallow lake in Florida, USA. *Hydrobiologia*, 581: 25-42.
- [66] Smolders, A., L. Lamers, E. Lucassen, G. Van Der Velde and J. Roelofs, 2006. Internal eutrophication: How it works and what to do about it – a review. *Chemistry and Ecology*, 22(2): 93-111.
- [67] Borell Löfstedt, C. and L. Bengtsson, 2008. The role of non-prevailing wind direction on resuspension and redistribution of sediments in a shallow lake. *Aquatic Sciences*, 70: 304 – 313.
- [68] Markensten, H. and D. Pierson, 2003. A dynamic model for flow and wind driven sediment resuspension in a shallow basin. *Hydrobiologia*, 494: 305-311.
- [69] de Vicente, I., V. Amores and L. Cruz-Pizarro, 2006. Instability of shallow lakes: A matter of the complexity of factors involved in sediment and water interaction? *Limnética*, 251(1-2): 253-270.
- [70] Echaniz, S. A., A. M. Vignatti and J. D. Segundo, 2011. Cambios en la diversidad y biomasa zooplanctónica durante una estación de crecimiento en un lago somero temporal hiposalino de La Pampa. *BioScriba*, 4(1): 1-12.
- [71] Paggi, J., *Cladocera (Anomopoda y Ctenopoda)*, 1998. Pp. 507-518. En: S. Coscarón & J. J. Morrone (eds), *Biodiversidad de Artrópodos Argentinos*. Ediciones Sur, La Plata.
- [72] Menu-Marque, S., J. Morrone and C. Locascio de Mitrovich, 2000. Distributional patterns of the south american species of Boeckella (Copepoda: Centropagidae): a track analysis. *Journal of Crustacean Biology*, 20(2): 262-272.
- [73] De los Ríos, P. 2005. Richness and distribution of zooplanktonic crustacean species in Chilean altiplanic and southern Patagonia ponds. *Polish Journal of Environmental Studies*, 14: 817-822.
- [74] Vignatti, A. M., S. A. Echaniz and M. C. Martín, 2007. El zooplancton de lagos someros de diferente salinidad y estado trófico en la región semiárida pampeana (La Pampa, Argentina). *Gayana*, 71(1): 38-48.
- [75] Pejler, B., 1995. Relation to habitat in rotifers. *Hydrobiologia*, 313/314: 267-278.
- [76] Echaniz, S. A., A. M. Vignatti and G. C. Cabrera, 2009. Características limnológicas de una laguna turbia orgánica de la provincia de La Pampa y variación estacional del zooplancton, *Biología Acuática*, 26: 71-82.